Mineral sequestration for CCS in Finland and abroad

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Abstract: The long-term storage of CO_2 using mineral sequestration is becoming increasingly interesting in many regions, especially where CO_2 underground sequestration is considered impossible or unfeasible. Despite the recognised and documented advantages of CO_2 mineral sequestration, twenty years of R&D work did not yet result in mature, economically viable technology that can be applied on a large scale. Lacking other CCS options while having access to large resources of suitable rock material, a route for carbonation of magnesium silicate mineral is currently being optimised in Finland. It involves the production of magnesium hydroxide, Mg(OH)₂ from the mineral followed by carbonation of this in a pressurised fluidised bed reactor. Although the Mg(OH)₂ production requires energy the consequent carbonation step is exothermic and the overall process could still be rendered energy neutral. Significant amounts of iron oxides are obtained as by-products. Carbonation levels of ~50% of several 100 μ m diameter Mg(OH)₂ particles were obtained within 10 minutes at pressures > 20 bar and temperatures up to 500°C. This paper reports on the latest developments of the work, addressing also process energy efficiency. Also, the large-scale application of this in Finland and at the locations of project partners abroad is briefly addressed.

Keywords: Carbon dioxide sequestration, Mineral carbonation

1. Introduction

The long-term storage of CO₂ using mineral sequestration is becoming of increased interest in many regions, especially where CO₂ underground sequestration is not possible or considered unfeasible. At many locations worldwide very large deposits of suitable mineral, usually magnesium silicates (serpentine, olivine) but sometime also calcium silicates (wollastonite) are available. Examples for this are Finland, East-coast Australia, Portugal and regions at the west coast of the USA and Canada. Despite the recognised and documented advantages of CO₂ mineral sequestration (very large capacity, no pos t-storage monitoring needed, exothermic overall process chemistry) the development work is still in the laboratory demonstration scale: twenty years of R&D work did not yet result in mature technology that can be applied on a large scale in an economically viable way [1,2,3]. Motivated by the slow deployment of large scale underground storage of CO₂ or simply the availability of large amounts of suitable mineral, progress on mineral sequestration is being steadily made and reported from an increasing number of research teams and projects worldwide. Also, increasingly realistic understanding of usable storage capacity for underground sequestration is changing the relative positioning of different CCS methods [4]. As a result, CO₂ mineral sequestration shows a clear trend towards scale-up and commercialisation, as is further illustrated by a significant number of patents awarded quite recently [5,6]. In addition, the issue of what to do with the solid product material has resulted in developments in CO₂ mineral sequestration towards both low value (land reclamation) and high value (pharmaceutics) applications.

Development work in Finland, where the exothermic carbonation chemistry is the reason for focussing on hi gh temperature, gas/solid carbonation at elevated pressures involves cooperation with a growing list of international partners, such as in the Baltic states Estonia and Lithuania but also in the Netherlands, Portugal, UK and more recently also Canada and Singapore. The route for carbonation of magnesium silicate mineral as currently being optimised at Åbo Akademi University (ÅA) in Finland involves the production of magnesium

hydroxide, Mg(OH)₂ from the mineral followed by carbonation of this in a pressurised fluidised bed (PFB) reactor. Although the Mg(OH)₂ production step requires energy the consequent carbonation step is exothermic and the overall process could still be rendered energy neutral (or even negative). This energy recovery distinguishes the method from other routes for CO₂ mineralisation. In addition, significant amounts of iron oxides are obtained as by-products [7]. Carbonation levels of ~50% of several 100 μ m diameter Mg(OH)₂ particles where obtained within 10 minutes at pressures > 20 bar and temperatures up to 500 °C [8-11]. The production of $Mg(OH)_2$ currently requires more heat than is generated by its carbonation, but nonetheless this route shows similar or better energy economics (0.9 - 1.2 vs. 1.0 - 2.3 kWh/kg CO₂ fixed) than the more straightforward route that is widely considered as "state of the art", i.e., direct mineral carbonation of superheated aqueous suspensions under high CO₂ pressure [12, 1 (p. 326)]. The route via Mg(OH)₂ also shows (much) faster carbonation kinetics than the conventional process, especially for larger particles - see below for more detail and results. One important benefit of the stepwise approach is that oxides of iron and calcium are obtained as separate by-products, such that magnesite (MgCO₃) is the unique carbonation product. The main mineral contaminant (iron oxide) thus extracted is sufficiently abundant to be of interest to the iron- and steelmaking industry [13].

This paper reports on the latest results of the development work where the carbonation of several rock types are compared, addressing the production of $Mg(OH)_2$ and the rate and final level of $Mg(OH)_2$ carbonation, and process energy efficiency. Also, the application of this for large-scale CO_2 mineral sequestration in Finland and at the locations of our project partners abroad will be addressed. This includes the use of the (by-) products of the process which can be used for land reclamation (as is an objective in Singapore), heat storage, or iron- and steelmaking. Finding such uses or markets for the solid products and by-products is essential.



Fig. 1. A schematic illustration of the mineral carbonation process under development at ÅA

2. Process description: serpentinite carbonation via $MgSO_4$ and $Mg(OH)_2$

The staged process under development at ÅA is schematically given in Fig. 1. As raw material, serpentinite rock rich in serpentine $(Mg_3Si_2O_5(OH)_4)$ is considered (being abundant in Finland), although most carbonation tests have been made using a commercial $Mg(OH)_2$ sample. Besides this, tests were made with magnesium silicate-based rock material from Lithuania, Portugal, Australia and other locations [14,15]. Table 1 lists selected composition data for some of the materials studied.

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Rock	MgO (% wt)	CaO (% wt)	$\begin{array}{c} \operatorname{Fe_2O_3} * \\ (\% \text{ wt}) \end{array}$	SiO ₂ (% wt)	$\begin{array}{c} Al_2O_3\\(\% \text{ wt})\end{array}$	Others (% wt)
Hitura, FI	38.1	0.5	14.8	47.6	10.0	6.2
Vammala, FI	14.5	5.6	12.5	49.5	8.8	9.1
Varena, LT	31.4	1.2	17.6	34.0	0.5	15.3
Braganca, PT	35.8	< 0.1	8.2	41.9	1.2	12.9
Great Serpentine Belt, AU	49.0	< 0.1	6.9	41.9	1.8	0.4

Table 1. Composition of some of the magnesium silicate rock samples being studied

* Calculated, presumably a mixture of FeO and Fe_2O_3 , i.e. Fe_3O_4 .

2.1. $Mg(OH)_2$ production

In the first process step, (preheated) serpentinite rock is thermally treated with ammonium sulphate (AS) at 400 - 500 °C and atmospheric pressure for 10 - 60 minutes. A significant amount of the magnesium, Mg, in the rock is thus converted to sulphate, MgSO₄, which is highly soluble in water. Unfortunately, MgSO₄ cannot be directly converted with CO₂ to MgCO₃, but in an aqueous solution it can be converted to $Mg(OH)_2$. After cooling, the solid from the reaction with AS is slurried in water, leaving behind unreacted mineral and insoluble reaction products, e.g., silica. The pH of the filtrate solution is raised to 8 - 9, precipitating iron and calcium (from the mineral, see Table 1) as FeOOH and Ca(OH)₂, respectively, while increasing the pH further to 10 - 11 precipitates Mg(OH)₂. For the Finnish Hitura mineral, the preferable conditions for extraction of Mg (and Fe) to MgSO₄ (and FeSO₄) are temperatures 400 - 440 °C, for 30 - 60 minutes at S/AS = 0.5 - 0.7 kg/kg, with 60 - 66 % extraction of Mg. Lower temperatures and longer reaction times give a higher (relative) extraction of iron. Ammonia vapour, NH₃, released during the thermal step is collected and used to give the necessary pH increases for precipitation. It is thereafter recovered for regeneration of the AS salt downstream, using heat from another process step. Nonetheless, the recovery of solid ammonium sulphate from the aqueous form incurs a not insubstantial energy penalty.

2.2. $Mg(OH)_2$ carbonation

The Mg(OH)₂ produced as described above is converted into MgCO₃ in a pressurised fluidised bed (PFB) reactor at pressures > 20 bar and temperatures 450 - 600 °C. Results on conversion levels obtained under varying conditions (temperature, pressure, water content of the gas, time, fluidisation velocity) are reported elsewhere [9-11] for both the synthetic, commercial Mg(OH)₂ material and Mg(OH)₂ produced from Finnish or Lithuanian serpentinites. A few tests were made under supercritical CO_2 conditions (pressure > 74 bar) which showed significantly lower conversion levels and rates, suggesting that little benefit should be expected from operating at such pressure levels. It was found that the Mg(OH)₂ materials produced from the serpentinites are much more reactive (as a result of a $\sim 10 \times$ larger specific surface of ~45 m²/g vs. ~5 m²/g), giving conversion levels of 50 % within 15 minutes for ~300 µm particles. The product gas from the carbonator is a hot, pressurised mixture of CO₂ and H₂O, the solids obtained will be partly recycled for further carbonation conversion. Unfortunately, although the carbonation reaction is rapid it levels off at a carbonation level of 50-55 % for the synthetic, commercial $Mg(OH)_2$ which may be the result of calcination of Mg(OH)₂ to MgO. However, it is noted that in order for Mg(OH)₂ to carbonate, dehydroxylation (i.e. calcination) has to occur. Apparently, carbonation takes place at a slower rate than dehydroxylation, resulting in a partially calcined and carbonated product. The amount of Mg(OH)₂, MgO and MgCO₃ in samples after test is plotted in Fig. 2 as a function of temperature for ~ twenty experiments (CO₂ pressure range 20 - 58 bar) with varying experiment time, fluidisation velocity, particle size, etc.



Fig. 2. Composition of Mg-species as a function of temperature for various experimental conditions using a PFB reactor [11].

2.3. Process (energy) efficiency

One of the features of CO₂ mineralisation using magnesium silicates is that the overall chemical reaction is exothermic. However, the direct carbonation of magnesium silicates is too slow, too energy demanding, or otherwise economically unviable, although work on improving the rate of processes based on pressurised aqueous solutions is still ongoing at several locations [3,6]. For the process route presented above an analysis was made of the heat requirement of the thermal treatment of Finnish Hitura (nickel mine tailing) serpentinite (see Table 1) with ammonium sulphate and the heat generated by carbonation of the resulting Mg(OH)₂. This showed that producing Mg(OH)₂, at 400 – 500 °C will require 4× more heat than what is obtained, at 450 - 550 °C, from carbonating it. Although ~1.2 MJ/kg CO₂ can be recovered as reaction heat the overall heat input requirements add up to 4 - 5 MJ/kg CO₂, consuming 3 - 4 ton rock per ton CO₂ [16]. An improved design using pinch analysis to optimise the heat exchanger network of the process, followed by process simulation with Aspen Plus[®] and exergy analysis (of the process heat input and outputs) reduces this to ~ 3 MJ/kg CO₂ heat input requirements while consuming \sim 3.1 ton rock per ton CO₂. The regeneration of ammonium sulphate (AS), which is obtained as aqueous solution after Mg(OH)₂ precipitation (while powdered, dry AS is used in the thermal treatment of serpentinite) puts a high energy penalty on the process [16].

It has been reported [17] that AS crystallisation from an aqueous solution can be accomplished at ~90 °C against a moderate heat input of ~120 kW/m³ (residence time 95 minutes in a 0.97 m³ DTB crystalliser). Nonetheless, a less energy consuming alternative for the AS recovery must be found and the solid/solid extraction must be improved, not only for energetic reasons, but also to recover more by-products thereby reducing the amount of solid residue. The rather high solubility of magnesium sulphate and ammonium sulphate in water should allow the use of minimal amounts of water in the precipitation steps towards Mg(OH)₂. Further improvement is obtained by optimising the different temperatures in the three aqueous precipitation steps. A variety of process refinements that lead to better energy efficiency, extraction from serpentinite, and AS recovery, was recently reported by Romão et al. [14] A recent study by Björklöf [18] applies mechanical vapour recompression (MVR) for the recovery of AS salt, making use of pinch analysis combined with chemical exergy analysis (in a spreadsheet calculation). The outcome of the study gives an energy penalty of 5.54 MJ/kg CO₂ fixed, expressed as exergy (using conservative data for the magnesium

sulphate solubility in water). Fig. 3 gives a so-called Grassmann diagram that clearly points out the exergy destruction in the various stages. Similar to Romão et al. [14,16] the energy penalty of the process (as exergy losses) was identified to arise primarily from 1) the AS recovery from dilute aqueous solutions, 2) the magnesium extraction using AS, and 3) the cooling of hot extraction products to aqueous solution temperature.



Fig. 3. A Grassmann diagram of the staged ÅA process. The gray triangles represent exergy destruction in the various process steps and the arrows represent exergy losses. At the far right, the products consist of the chemical exergies of the products and the recoverable heat from the carbonation step. [18]

2.4. Magnesium extraction and carbonation efficiency, ammonium sulphate recovery

Besides carbonation efficiencies for Mg(OH)₂ in the PFB levelling off at 50-55 % (for a synthetic, commercial sample) also the extraction of magnesium from serpentinite needs improvement, with extraction efficiencies obtained so far seldom exceeding 60% of the Mg content of the rock. For this, development work commences at ÅA, aiming at using a rotary kiln for the magnesium extraction, instead of using a fixed bed ("heap") for the conversion because this requires higher temperatures than necessary to compensate for heat and mass transfer limitations. The excess temperature leads to irreversible loss of the AS salt as SO₂ and N₂O. Detailed chemical reaction and solid product analysis suggest that temperatures should not exceed 400 °C. The possibility of losses of AS, e.g., occluded within the solid residue of unreacted serpentinite, silica etc., or in the Mg(OH)₂ fed to the carbonator (probably as NH₄⁺ & SO₄²⁻ ions) was addressed by Björklöf [18]. Analysis showed that the solid residue contained < 0.1 %-wt nitrogen and < 0.9 %-wt sulphur which corresponds to ~ 0.15 % and ~ 1.5 %, respectively, of the incoming AS. For the Mg(OH)₂ the nitrogen – and sulphur contents were 0.1 %-wt and 0.8 %-wt, respectively.

3. Implementation of the results in Finland and abroad

Finland, like many countries in the EU, has commitments with respect to greenhouse gas emissions under the Kyoto Protocol and a continuation of the use of fossil fuels may be difficult without also implementing a CCS method. In Finland, the following schemes can be considered – see also Fig. 4 for industry sector integration:

CO₂ from large-scale producers in Central / Northern Finland can be fixed using the vast resources of serpentinite-containing rock (estimated CCS capacity 2.5 – 3.5 Gt CO₂ [19]). One example is Ruukki's iron- and steelmaking plant at Raahe, which is ~110 km from the nickel mine at Hitura where large amounts of mine tailings are deposited. At the same time, large amounts of iron oxide by-products will be obtained from the rock, ready for use at the iron- and steelmaking plant. And, the slag by-products, most importantly steel

converter slag, may be carbonated to yield valuable calcium carbonates of PCC (precipitated calcium carbonate) quality [20].



Fig. 4. Integrated processing for magnesium silicate carbonation, iron- and steelmaking and steel slag carbonation [13]

• With most CO₂ produced in Southern Finland, lower-grade minerals than found in Central/Northern Finland must be made use of. At several locations in South / South-west Finland rock with an MgO content of 10-15 %-wt were found, offering potential for source-sink combinations such as

* Vammala serpentinite for use at the coal-fired plant at Meri-Pori (distance ~ 90 km)

* Suomusjärvi serpentinite for use at a lime kiln in Parainen (distance ~ 90 km) or at a power plant in Naantali (distance ~ 95 km)

Also abroad the technology can be made use of, for example our cooperation with:

- Lithuania, where significant and suitable (although maybe located somewhat deep) serpentinite were found in the Varena region in the South-east of the country [15]
- In Portugal, suitable rock was located at several locations within the country [14], offering good opportunities for CO₂ mineralisation while also the option of (on-shore) underground sequestration is being investigated [21]
- Singapore, where CCS is combined with land reclamation during the next decade, using rock material that is imported from the region, for example from Australia [22]

In most of the cases, except the last mentioned obviously, CO_2 would be transported to the mineral site. On the other hand, like fossil fuels, metal ores or other raw material also rock transport can be feasible, with the advantage that the carbonation process can operate directly on the CO₂-containing gases. This removes the CO₂ capture step from the CCS chain.

4. Conclusions

The performance and efficiency (with respect to energy and chemicals recovery) of a staged process for serpentinite carbonation as under development at ÅA was described and assessed. It involves the production of magnesium hydroxide, $Mg(OH)_2$ from the mineral using ammonium sulphate, AS (which is later recovered) followed by carbonation of this in a pressurised fluidised bed (PFB) reactor. The process can be considered to be a variety of a process route patented by Pundsack in 1967 [23], which is based on extraction of magnesium from serpentine using an aqueous solution of ammonium bisulphate (ABS). The ÅA route for serpentinite carbonation, instead, uses AS in a high temperature step for magnesium

extraction that proceeds much faster than that with ABS in an aqueous solution (<< 1 h vs. >> 1 h) whilst also the reaction heat release from the carbonation is taken advantage of in the gas/solid reactor. However, recovery of solid ammonium sulphate from the aqueous form incurs a not insubstantial energy penalty. Thus, there seems to be a need to develop an alternative route to Mg(OH)₂ that bypasses the aqueous stage, and/or a route in which MgSO₄ is carbonated directly, e.g., using ammonium (bi)carbonate produced from CO₂ absorption in (aqueous) ammonia in an upstream scrubbing stage. Ammonium sulphate is a cheap and abundant reagent for extracting magnesium from serpentinite, but its performance must be evaluated under milder conditions as predicted by thermodynamics. The other technical issue linked with its use is the containment of NH₃ in the system.

Ammonium salts, i.e., (bi-)carbonate and (bi-) sulphate, could play an important role for extracting magnesium from rock material and as reactants produced from scrubbing CO_2 from process gases with chilled aqueous ammonia solutions – see for example [24]. An important benefit of the CO_2 mineral sequestration routes that look most promising for scaleup and large-scale application is that the expensive and potentially problematic CO_2 capture stage can be removed from the CCS chain. This is one of the drivers of current interest for CO_2 mineral sequestration. It is also considered for the ÅA process route, although the gas/solid carbonation will require CO_2 partial pressures > 20 ba r. On the other hand, scrubbing CO_2 from a power plant flue gas will introduce water and other species (and potential contaminants) into the process loop, eventually contaminating the sorbent.

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