Model selection for waste conversion efficiency and energy demands in a pilot for large-scale larvae treatment

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Abstract
Implementations of conversion systems for segregated food waste into larval biomass are reliant on stability of production environment and predictable outcomes. However, the knowledge is currently lacking for large scale implementation modelling allowing to adjust the process for reaching a stable production quality and predicting the treatment capacity and output product quantity. This study contributed to the development of such knowledge and investigated models for prediction of larvae mass gain in the conversion process, food waste degradation due to larval and microbial activity, heat production based on metabolic activity in the conversion process. These models were used to evaluate the amount of heat produced and compared to the demand for water removal for achieving a minimal total solids content (50%) in the treatment residue for easy larvae separation by sieving. Based on these models’ predictions and measurements of the conversion efficiency it was established that, assuming no heat loss from the system, the heat generated by the process was sufficient to achieve a desirable total solids content in the residue after 14 days of treatment. An exponential heat production from waste degradation can be expected, and thus, for wetter food waste, the treatment period can be expanded to achieve the desired residue total solids content.

Keywords: Black soldier fly, degradation, conversion, sieving, food waste, high water content.

1 Introduction
Insect rearing for feed production is becoming more enticing in strive to find sustainable and locally produced feed proteins (Kim et al., 2019). Still, the trend globally is that consumption of insects used as feed and food is decreasing, although a recent estimation suggests that some 2.5 billion people occasionally eat insects, mostly in Asia, Africa and Central – South America (Van Huis et al., 2013). Historically, in Europe, the use has been quite low. However, the search for sustainable protein is driving the development of larger facilities producing insects (Makkar, 2017). A great deal of attention has been given to the black soldier fly (Hermetia illucens) in particular due to its potential for use in organic waste treatment (Čičková et al., 2015). A step-by-step guide for such implementation was recently presented by Dortmans et al. (2017). However, the majority of scientific investigations have so far focused on treatment efficiency in treating different organic waste fractions, and very few studies focus on large-scale production optimization (Mertenat et al., 2019). Other studies by Leong et al. (2016) and Li et al. (2011) investigated bio diesel production potential, while Pastor et al. (2015) and (Spranghers et al., 2017) have studied bottlenecks in conversion processes.

The success of large-scale insect biomass production system implementation is reliant on stability of production environment with predictable output quantities and quality of larvae and residue (Čičková et al., 2015). Such a system can be easily scaled to the expected amount of substrate and appropriate adjustments can be made with known effects on outcomes depending on the amounts of products needed or substrate to be treated. Such stability has to be achieved by insect farmers at European and global level in order to meet future demands for protein (IPIFF, 2018). The use of organic waste as a source for production of larvae with intent for animal feed or biodiesel production has been evaluated in several studies (Leong et al., 2016; Li et al., 2011; Pastor et al., 2015; Spranghers et al., 2017). The grown black soldier fly (BSF) larvae are high in fat, containing 20-50% of fat on a dry matter basis. Protein content is typically in the range of 35-42% on a dry matter basis (Ewald, 2019; Lalander et al., 2019). Such composition of larval biomass represents a great source of protein and fat for production of animal feed.

For large scale implementation, several parameters are important to establish a stable production. Such parameters include the dimensioning of the number of fly larvae per amount of treated substrate, process temperature and retention time in treatment (Diener et al., 2011; Liu et al., 2017; Paz et al., 2015; Tomberlin et al., 2009). Another important parameter often discussed is achieving a moisture content in the treatment residue that allows dry separation of larvae after the treatment. Cheng et al. (2017) has investigated how larvae
separation is affected by moisture content in the residue. The initial tests performed at the pilot plant at Eskilstuna Strängnäs Energi och Miljö collaborating in the current project have shown that total solids content of the residue above 40-45% allows an easy larvae separation through sieving.

The aim of this study was to evaluate the technical aspects of larvae growth, ventilation and drying in a pilot sized BSF treatment system treating segregated household food waste and to develop mathematical models to describe the treatment process and its energy efficiency.

2 Material and methods

This section contains details about the study set up, materials and methods used, measurement techniques and assumptions.

2.1 Description of the larvae production facility in Eskilstuna

The measurements were performed at the fly larvae conversion pilot facility at Eskilstuna Strängnäs Energi och Miljö, while process efficiency evaluations were performed both at the pilot facility and at the Swedish University of Agricultural Sciences (SLU).

2.1.1 Treatment set-up

The treatment process began when approximately 15 000 seven-day-old larvae were added to a treatment box containing the first feeding portion of 5 kg waste (Figure 1A). The treatment boxes had a surface area of 2400 cm² (Figure 1). Treatment boxes were stacked by 8, and four such stacks were started approximately once per week. The boxes were stacked in order to increase treatment volume. The stacks were placed in a treatment room which was heated using residual heat at the Eskilstuna waste management facility. Three sets of stacks (4 in each) were evaluated. The sets were named T20, T11 and T5 indicating how many days they were in the treatment on the day when temperature and relative humidity (RH) were measured. Four boxes were sampled in each set.

2.1.2 Materials

The young larvae were produced at SLU from a continuously run BSF colony since 2015. The seven-day-old larvae were pre-weighed to contain approximately 15 000 based on a smaller sample that was determined by enumeration and a gravimetric estimation of a single larvae weight.

The segregated household food waste, which was collected in plastic bags and went through optical sorting process, subsequently was minced in an extruder-grinder that also separated out plastic and other fractions not passing through the mesh.

Figure 1. Depiction of larval growth and visual material changes during treatment process a) at start of treatment, b) after 5 days of treatment, c) after 11 days of treatment and d) after 20 days of treatment.

<table>
<thead>
<tr>
<th>Stack set</th>
<th>Treatment days</th>
</tr>
</thead>
<tbody>
<tr>
<td>T20</td>
<td>ST 1 2 3 4 5 6 7 8 9 10 11 12 13 14 15 16 17 18 19 20</td>
</tr>
<tr>
<td>T11</td>
<td>ST 1 2 3 4 5 6 7 8 9 10 11</td>
</tr>
<tr>
<td>T5</td>
<td>ST 1 2 3 4 5</td>
</tr>
</tbody>
</table>

Table 1. Treatment schedule showing the age of stacks with treatment boxes in days from start (ST) to sampling day (highlighted in green). First feeding was provided at start (highlighted orange) and additional feedings were given within 3-6 days (highlighted blue).
compared to ensure reliability of measurements (n=9 for temperature and n=4 for surface humidity).

2.2 Modelling

Microbial waste degradation was modeled according to a first order rate expression for microbial oxidation of degradable volatile solids highlighted by Hamelers (2004) and developed by Keener et al. (1992) (1). Similar models based on rate constants, amount of degradable material and environmental factors such as process temperature have been discussed by others e.g. Haug (1993).

\[
\frac{dm}{dt} = -k(x_1, x_2, \ldots, x_n) \cdot [m - m_e]
\]

Where \( m \) is the compost mass, \( t \) is time, \( k \) is the composting process rate constant, \( x_i \) are the environmental factors e.g. process moisture and temperature and \( m_e \) is the non-degradable mass of compost at the end of the process.

Larval waste reduction was modelled based on several variables, including bioconversion ratio (BCR) representing the amount of volatile solids (VS) in substrate that has been converted to larval biomass (2).

\[
BCR_{VS} = \frac{V_{S_{larvae}}}{V_{S_{substrate}}}
\]

The mass of substrate lost in the process is expressed through a reduction (RED) variable which can be calculated on volatile solids basis, total solids (TS) basis and wet weight (WW) basis. The latter accounts for all the losses in the process including evaporation and respiration (3).

\[
RED_{WW} = 1 - \frac{WW_{larvae} + WW_{residue}}{WW_{substrate}}
\]

The larval metabolic activity, along with that performed by microbes, results in respiration, energy release and volatile solids reduction in the substrate. Typically such activity can be associated to the body mass and thus the substrate VS reduction by larvae can be expressed through an adapted model proposed by Wotton (1978) where respiration rate (measured as C loss from substrate through CO₂ emission) is assumed to be proportional to VS loss (4).

\[
\log \frac{dV_{S_{loss}}}{dt} = a + b \cdot \log M_{larva}
\]

Where \( V_{S_{loss}} \) represents the amount of substrate VS being lost due to larval metabolic activity over time \( t \), \( M_{larva} \) is the individual larvae body mass at time \( t \) in the process, while parameters \( a \) and \( b \) represent the initial VS loss rate and the influence degree of the body mass on VS loss rate, respectively. An additional pathway for VS loss is associated to substrate assimilation by the larvae and can be calculated based on BCR_{VS}.

2.3 Assumptions

Several parameters used in modelling and evaluations of system productivity were estimated based on previous analyses, studies and published sources. Eskilstuna food waste moisture content has been evaluated in several studies including a study by Johannesdottir (2017) and Lindberg (2018) and the TS range reported was 15%–23%. Based on these results, communication with Eskilstuna facility operators and measurements performed in the lab (data not shown), it was assumed that the typical TS content of the slurry was 16% ± 2%. Food waste from Eskilstuna have been demonstrated to have a very good BCR_{TS} the values reported range between 24 and 32% (Johannesdottir, 2017; Lindberg, 2018) and in the current system investigation a value of BCR_{TS} of 30% was assumed. Reduction of substrate VS depends on multiple parameters and system dimensioning; however, an assumption is necessary in order to estimate the amount of water needed to be removed in the treatment process. Although a reduction in the range of 56 - 65% (Johannesdottir, 2017) have been reported for Eskilstuna food waste, a lower RED_{TS} of 50% was assumed here in order to not overestimate the amount of energy released from the process typically associated with reduction of substrate.

Based on these values the mass balance of the process can be presented. Per ton of food waste processed, the substrate TS represent 140-180 kg, which assuming a BCR_{TS} of 30% and a RED_{TS} of 50%, results in production of 43-55 kg TS of larval biomass and 72-93 kg TS of treatment residue. Based on these assumptions, the demand to obtain the residue with 50% TS to allow sieving, the assumption that larvae have TS of 35% (Johannesdottir, 2017) and the overall mass balance, 709-625 kg of initial substrate water has to be removed and 25-32 kg of initial TS are respired.

The amount of energy released in the degradation process was estimated based on heat of combustion of biodegradable VS. Mason (2009) calculated that for food waste 5.4 kWh of energy can be released per kg VS degraded. Based on this assumption the amount of heat released was calculated taking into account the VS content of the food waste slurry (He et al., 2013).

The incoming air to the facility was warmed up to approximately 30 °C and it was assumed it had 15 RH% at that temperature based on measurements performed earlier (data not shown). The outgoing air was assumed to be 30 °C and have an RH of 95%. Such high saturation level was not achieved in a previous study performed in controlled lab conditions (Johannesdottir, 2017) when a saturation of 61 RH% was observed in outgoing air. Nevertheless, by adjusting the ventilation rates, nearly full saturation of air can be achieved (Bach et al., 1987), and thus this assumption was used in the current evaluation (Kubilay and Kucskka, 2018). Based on these assumptions a water carrying capacity of air was calculated using steam tables (Mörtstedt and
Hellsten, 1976). It was estimated that 0.022 kg water can be removed per kg air supplied by the ventilation.

3 Results and discussion

3.1 Treatment process

The treatment process has been running for 5, 11 and 20 d on the day of moisture sampling. The treatment running for 5 d (T5 in Table 1, Figure 1B) had larvae that were approximately 10 mm long (visual estimation) but very thin. In the treatment T11, running for 11 days (Figure 1C) the larvae were larger, visually estimated to be approximately 15 mm long, but smaller in diameter (by ca 2.5-3 mm) than the ones in treatment for 20 days (Figure 1D), which were estimated to have a diameter around 3.5 - 4 mm and were approximately 20 mm long.

The moisture content in the material was relatively high after 11 days of treatment, and as seen in the picture in Figure 1B, the treatment surface appeared wet. Figure 1D shows the larvae that had treated the material for 20 days. The waste was dried to a TS content of approximately or lower than 50% due to biological activity generating heat that was driving evaporation together with ventilation.

The facility had a treatment capacity for 1440 kg of material at any given time, however as the material was added in portions, on the sampling day there was 960 kg of material being treated. Only stacks in the treatment T20 reached the full capacity as they were fed 3 times (Table 1). The average capacity for the facility considering the feeding frequency and treatment duration of 21 days was 69 kg waste/d, however the capacity could be increased to 103 kg waste/d if a shorter treatment time was used. The available space was not fully utilized, and the design goal was to reach a treatment capacity of 1000 kg/d, which would be possible if the process was run during 14 days per batch and all potential treatment volume was used.

During this evaluation, the pilot fly larvae treatment facility in Eskilstuna used stacked treatment units (boxes) each treating 15 kg of food waste. Each unit after 14 d of treatment produced approximately 2.4 kg of larvae and 2.5 kg residue (based on initial TS and VS evaluations performed in the lab and assuming a BCR$_{TS}$ of 30% and RED$_{TS}$ of 50%, data not shown). The estimation of the process performed in the lab (based on TS and VS measurements of substrate and treatment products, data not shown) on dry solids basis gave 1.0 ± 0.2 kg larvae and 1.0 ± 0.1 kg residue and a mass loss due to metabolism of 0.5 ± 0.1 kg TS (variations assumed based on expected variation in moisture content in the incoming material). On the measurement day a total of 11 treatment units were sampled with respect to temperature and relative humidity (T5, T11 and T20) from the three treatments sets (Figure 2).

The difference in temperature below the surface (in material) and at the material surface was approximately 0.5 °C for T5, while it was 2 °C for T11 and 2.7 °C for T20 (c.f. standard deviations Figure 2). This shows that the longer treatment process was, the higher heat generation was observed, increasing the temperature inside the material and causing larger difference in surface temperature.

3.2 Modelling

Several processes can be modeled in this system implementation that should follow standard models, including larval and microbial growth, energy generation in the process and water loss from material due to evaporation and ventilation. The general microbial and larval waste degradation models are presented in the methods section (1 and 4). The larval degradation model uses larval mass as a parameter for estimation of degradation (4). It is thus important to evaluate the larvae growth and develop a model for its characterization.

![Figure 2](attachment:image.png)

Figure 2. Graphical representation of a) %RH and b) temperature (T) on the measurement day in different treatments, whiskers (where available) show standard deviation and numbers inside bars show n. Measurement of T b) and %RH a) in boxes using pyrometer (T1) from Fluke and T (T2) and RH% using Testo 400.

3.2.1 Larvae growth and material reduction

The BSF larvae growth have been evaluated in multiple studies (Diener et al., 2011; Lander et al., 2019)
showing that larvae grow nearly linearly, but there is a slightly slower growth rate at the beginning of the process. This suggests that an exponential growth model could fit well to describe the first period of larvae growth during the first 10-14 days (5).

\[ x_{d+1} = x_d + a \cdot t^b \]  

Were \( x_{d+1} \) is the weight of larvae at next time step (next day, \( d+1 \)), \( x_d \) the weight at current time step (current day, \( d \)), \( a \) is a constant expressing the larvae mass gain rate, \( t \) is the accumulated time, in hours since the growth started and \( b \) is another constant expressing the larvae mass gain rate change with time. Constants \( a \) and \( b \) include environmental impacts of the system on larval growth. When the equation was fitted to achieve the expected larvae size in treatment over 10-14 days based on assumptions described in section 1.1.3, the constants \( a \) and \( b \) had values of 0.12 and 0.8, respectively, while \( a \) value became 0.15 when the treatment duration was set to 21 days. In this case assuming the treatment started with 160 kg TS in waste and after 14 days produced 49 kg larvae TS. The conversion efficiency (BCR\(_{TS}\)) in this case became \((49\times100\%)/160 = 30.6 \%\). Such a BCR\(_{TS}\) value represents a very good conversion ratio since the normal range is between 25 – 36 \% for different types of substrates (Lalander et al., 2019). Such a BCR range for food waste would result in 41-59 kg larvae from the original 0.25 TS kg young larvae and 160 kg TS waste.

It was observed by Lalander et al. (2019) that the BSF larvae grow mostly during the first two weeks of treatment, but then the additional weight gain is small during the third week, or even a weight reduction was observed. Such a development is associated with BSF life cycle: if the larvae have reached their target larval stage and become pre-pupae, they no longer feed and thus only consume stored fat and protein necessary for their metabolism, movement and metamorphosis. A further investigation of larval growth based on a study by Lalander et al. (2019), showed that the growth rate is slow in the beginning of treatment of food waste, but then accelerates and peaks during days 4 to 9. After this, the growth rate decreases, and after day 12 (for most substrates) the larval mass increase stops (Figure 3). When further increase in larvae mass stopped, but larvae have not yet reached the pre-pupae stage (days 12-14) the waste was degraded due to metabolism, but it did not give increased growth resulting in unchanged BCR and increased RED. After 3 weeks in treatment, the larvae start pupating, which causes an additional weight loss as they have stopped eating already in the pre-pupae stage. The peak weight gain and the duration of weight gain period was observed to differ between the substrates, but for food waste, the peak could be expected around day 7 (based on model projection and data from Lalander et al., 2019) under optimal conditions (Figure 3). This means the exponential model (5) well describes the beginning of the process, but to cover the whole treatment time a cumulative distribution function fits better and matches with the data points presented in Lalander (2019).

When comparing the actual weight curve to an exponential curve it was found that it fitted fairly well the first 10 days, but after that the larval growth stopped as the larvae reached their maximum weight (Figure 3). From a production perspective thus, a shorter harvest after approximately 10 days should be evaluated to maximize the output of larval biomass. Such shorter treatment can result in insufficient water removal from material (treatment residue) which needs to be sufficiently dry to enable separation of larvae (Cheng et al., 2017). In this pilot set-up, the mixture was still quite wet after 10 d and it would not have been possible to efficiently separate the larvae from the treatment residue. If the larvae are left in the material for longer, their and microbial metabolisms further reduce the substrate, which can result in additional water evaporation due to associated heat generation. Such reduction can be modeled using equations 1 and 4 described earlier. According to these models and the model for larval growth (5), the larvae and microbial metabolic rates increase exponentially and thus the energy release from metabolism also increases exponentially during exponential larval growth.

In the beginning of treatment, the bioconversion rate is low which means that heat generation is low, and the food waste slurry is still very high in water content. In parallel with larval digestion of food waste a microbial process also takes place. A first order exponential or linear model for aerobic food waste degradation can describe such a microbial process well (Hamelers, 2004). Thus, in the beginning of the treatment due to lag in larvae growth the decrease in moisture content was marginal, and likely mainly due to microbial activity. As the larval mass increased, the larval activity also

![Graphical (dimensionless) representation of larvae weight change as a function of time for 1 ton food waste with 16% TS fed during 14 days and an associated CO2 emission rate based on VS loss due to larval metabolism (4).](image-url)
increased, resulting in a substantial heat release and evaporation of water. If the treatment is allowed to continue beyond the typical 14 d until 21 d, the larva first become prepupae and then turn into pupae, meaning they no longer degrade the substrate. However the microbial waste degradation continues and provides additional heat necessary for water evaporation from the residue.

The amount of waste that was converted into heat can be calculated based on oxidation reaction of the average food waste formula, resulting in production of energy and of CO₂ + H₂O. The amount of food waste oxidized was estimated based on degradation rate, \( \frac{dm_{\text{waste}}}{dt} \) and was calculated from multiplying the mass of larvae, \( m_l \), by the degradation per kg larvae for each day, \( A_i \), and the same for microorganisms (\( m_{\text{mo}} \), \( A_{\text{mo}} \)). As the accumulated larval and microbial biomass increases, the absolute degradation per day also increases (6).

\[
\frac{dm_{\text{waste}}}{dt} = m_{\text{waste}} \times (m_l \times A_i + m_{\text{mo}} \times A_{\text{mo}}) \tag{6}
\]

The accumulated heat release from TS degradation can also be presented as a function of time. Assuming 0.25 kg TS of young larvae was originally added to 160 kg TS waste and using an assumption that 5.4 kWh is released per kg TS respired (Mason, 2009), the total energy release can be estimated. Based on the assumptions made earlier in methods section 1.1.3, an energy release can be estimated. Based on the results obtained from the pilot scale treatment facility in this study, the heat generated by the waste reduction during the degradation process for water evaporation mostly relying on heat release associated with degradation process for water evaporation (Johannesdottir, 2017; Kubilay and Kucska, 2018).

### 3.2.2 Evaporation of water to dry the waste

The waste slurry used in the process in this pilot set-up was relatively high in water content (~16% DM), assuming that the facility would process 1 ton waste per day, 160 kg TS/d would also bring 840 kg water/d. To separate larvae from the residue, the treatment residue should reach around 50% TS (Cheng et al., 2017). Taking into account the material reduction of 50%, 72 - 93 kg water would be remaining in the residue after 14 days of treatment, equivalent to 625 - 709 kg water needed to be evaporated per day. The heat released by the larvae estimated above was 449 kWh during the first 10 days but 807 kWh after 14 days. For removal of one kg water it is required 0.628 kWh of heat for evaporation. Assuming that only the heat evolved from the degradation of the waste would be used for evaporation and there was no other heat loss, the total amount of water evaporated could be 449/0.628 = 715 kg water. This roughly corresponds to an original moisture content of 16 % TS. If more water needs to be evaporated, additional heat has to be supplied (from the ventilation air), or the process should be allowed to continue for longer. Two extra days would yield additional 179 kWh, which would be enough to achieve desired residue TS if the original TS of the waste slurry was 13.8%.

The BSF larvae generally move quite actively, the amount of heat generated due to this motion is also expected to be proportional to the larvae size up until the pre-pupae stage, when larvae start behaving differently. This makes the selection of optimal time for harvest relevant for optimizing the protein production and total yield, while avoiding the unnecessary losses due to metabolic activity not resulting in further larval mass gain. The harvest period should be extended by approximately one more week (beyond the usual 14 days), if pupae are required for e.g. fly production to maintain a BSF colony and produce eggs. This was the case for the pilot set-up evaluated in this study.

In case the generated metabolic heat is insufficient for evaporation, the heat energy from heated ventilation air could be used to remove excess water from the substrate. Assuming the incoming ventilation air had a temperature of 30 °C with a RH of 15% and the outgoing air was saturated to a RH of 95% at 30 °C, 0.022 kg H₂O could be removed per kg air. To remove a total of 715 kg water the demand at such conditions would be 32 500 kg air which would be equivalent to ventilation rate of 85 m³/h. The ventilation fan at Eskilstuna pilot facility was estimated to provide airflow over 700 m³/h suggesting a quicker than observed water removal would be expected. However, the ventilation air was not directed in a way to maximize its saturation and thus likely contributed to evaporation to a smaller degree.

Previous studies on a similar system also highlighted that a better air saturation would improve water removal rate from such fly larvae conversion system that was mostly relying on heat release associated with degradation process for water evaporation (Johannesdottir, 2017; Kubilay and Kucska, 2018).

### 4 Conclusions

Based on the results obtained from the pilot scale treatment facility and the experiments performed in laboratory environment it was possible to model the black soldier fly larvae development and the associated waste degradation rate attributed to both larval and microbial metabolism. Based on these models’ predictions and measurements of the conversion efficiency it was established that, assuming no heat loss from the system, the heat generated by the process was sufficient to achieve a desirable total solids content in the residue after 14 days of treatment. An exponential heat production from waste degradation can be expected, and thus, for wetter food waste, the treatment period can be expanded to achieve the desired residue total solids content.

Using the models presented in this study and based on environmental properties of the treatment and waste type and moisture content, it is possible to project if the amount of heat generated by the waste reduction during the treatment is going to be sufficient to remove the amount of water to achieve a sievable residue (reaching
total solids content of 50%). After the necessary calibration of the models for each particular system, they can be used as tools for deciding the duration of treatment and the amount of heat supplied by the ventilation.

Acknowledgements

We thank Eskilstuna Strängnäs Energi och miljö for their support in implementation of the pilot project, Naturvårdsverket and Swedish University of Agricultural Sciences (SLU) for the financial support of this project.

References


